

Derivative Spectrophotometric determination of Mercury (II) using 3,4-Dihydroxybenzaldehyde Thiosemicarbazone (DHB TSC) in presence of micelle medium

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Abstract: A rapid and sensitive method has been developed for the determination of Mercury(II) based on complexation reaction between the metal ion and 3,4,-dihydroxybenzaldehydethiosemicarbazone (DHB TSC) in the presence of non-ionic surfactant Tween-80. The important parameters affecting the analytical procedure were optimized. Absorption maximum for a ternary complex was noted at 375 nm. The reaction was found to be rapid at room temperature and absorbance remained constant for more than 24h. The method obeys Beer's law in the range 5.01 to 50.14ng/ml. The apparent molar absorptivity of $3.25 \times 10^5 \text{ L mol}^{-1} \text{ cm}^{-1}$ and Sandell's sensitivity 0.06ng/ml. The effect of foreign ions was tested by taking a constant concentration of metal ion and determining its concentration in the presence of ≥ 100 folds in excess of foreign ions. The method was successfully used in the determination of Mercury(II) in Sewage waste and Spiked water samples.

I. Introduction

Mercury is a serious environmental pollutant with toxic effects in all living organisms[1]. Its effects on the immune system is potentially harmful, possibly contributing to diseases such as leukemia[2]. It is usually present in natural waters at trace levels[3]. The lakes, rivers and coastal waters in the vicinity of industries that utilize mercury in production are important indicators of environmental pollution. The main species of mercury in natural waters to be identified and determined are inorganic mercury(Hg^{+2}) and methyl mercury (CH_3Hg^+). Recent reports estimate total mercury concentration in natural waters to range from 0.2 to 100 ng L⁻¹, while methyl mercury levels are much lower (approximately 0.05 ngL⁻¹)[4]. However, mercury in contaminated environmental materials may exist at levels of monograms per liter. The development of analytical methods for the determination of mercury is still a challenge. The determination of low concentration of mercury is a vital task. Therefore, considerable efforts and progress have been carried out to develop accurate, low cost and reliable methods for mercury determination in contaminated samples without any complicated processing steps[5]. The most common techniques the pre-concentration should be performed to determine ultra trace Hg(II). However these procedures could reduce the accuracy of the determination and require, a longer time in handling of sample. Among these pre-concentration techniques, cold vapor atomic absorption spectrometry has widely been used [6-8]. X-ray fluorescence spectrometry (XRF)[9], atomic fluorescence spectrometry (AFS)[10,11] and spectrophotometer[12-15] have been used to determine Hg(II) at trace levels. Each of the above techniques has its own merits, but each also shares some of the common problems of poor reproducibility and limited sample adaptability. Among these techniques, visible absorption spectrophotometer represents the most convenient technique because of the availability of the instrumentation, simplicity speed precision accuracy and low cost.

A series of chromogenic reagents has been reported for mercury (II) determination in different samples [16-20]. Most of these methods are suffered from the lack of sensitivity due to the significant interference of the excess of chromogenic reagent with the analyte at Wavelength. This problem was solved by employing soft Lewis base such as sulfur ligands. Thiosemicarbazone have been used for spectrophotometric determination because of their good selectivity and sensitivity, although for conventional spectrophotometric analysis in aqueous solutions, the low solubility of these thiosemicarbazone compound and their complex is a significant drawback. This drawback can be overcome by adding organic solvent (or) surfactants.

The recent literature on the analytical Applications of the entitled reagent 3,4 - dihydroxybenzaldehydethiosemicarbazone abbreviated as (3,4 -DHB TSC) has revealed no study on the use of reagent for mercury(II) determination.

Therefore, the goals of the present manuscript are focused on the synthesis and spectroscopic characterization (UV-Vis IR and ¹H NMR, Mass) of the DHB TSC reagent. Moreover the stoichiometry of the formed mercury (II)-DHB TSC chelate was elucidated in an attempt to develop an accurate method for the analysis of mercury(II) in different water and sewage samples.

Recently enhanced sensitivity in spectrophotometer was achieved by utilizing the ability of certain surfactants to sensitize the binary complexes of the metal ion with chromogenic ligands[21,22] sensitization are result of the replacement of acidic protons of the liganded dye molecule by surfactant[23] (or) adsorption of the metalreagent complex on the micelles of the surfactant [24] cationic, anionic and nonionic surfactants are often used to sensitize the metallochromic indicators . In this context, updated surfacactant – sensitized reaction have recently been developed in sepectrophotometr[25-29].

The present study was hence planned to suggest a very simple and reasonably good method for determination of mercury ions at low concentration, using the reported reagent. DHBTS as binary complex and sensitizing the reagent with Tween 80 as ternary complex using spectrophotometer which is still frequently used because of its low cost and simplicity.

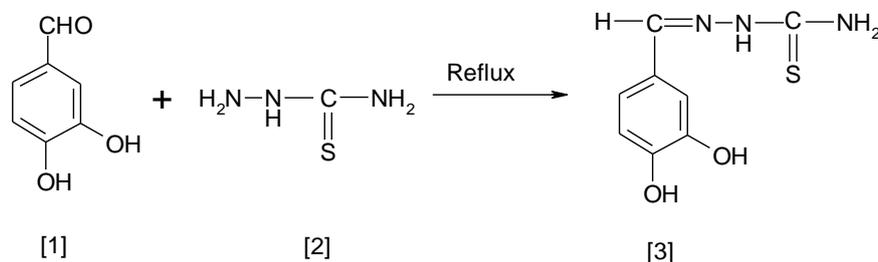
II. Experimental

2.1 Apparatus

The Absorbance and pH measurements were made on a Shimadzu UV-visible spectrophotometer (Model UV -160A) fitted with 1cm Quartz cells and Philips digital pH meter (model L1 613 respectively.)

2.2 Reagents and solutions

2.2.1 Preparation of 3,4-di hydroxybenzaldehydethiosemicarbazone(DHBTS) The Reagent (DHBTS) is prepared by the sah and Daniels procedure. 7 g of 3,4-dihydroxybenzaldehyde(1) was dissolved in 100 mL of methanol and 4.6 g of thiosemicarbazide(2) were dissolved in hot water. The two solutions were taken in 250 mL Erlymeyer flask and refluxed for 3 hrs. pale yellow colored crystals were separated out on cooling. These crystals were collected by filtration and washed several times with hot water and 50 percent methanol. The product(3,4-dihydroxybenzaldehydethiosemicarbazone)(3) (m.p 221-223⁰c) was recrystallisedtwice from boiling methanol and dried in vacuum.



All chemicals used were of analytical reagents grade or the highest purity available. Double distilled deionized water, which is non-absorbant under ultraviolet radiation, was used throughout. Glass vassals were cleaned by soaking in acidified solutions of KMnO_4 or $\text{K}_2\text{Cr}_2\text{O}_7$, followed by washing with concentrated HNO_3 and rised several times with deionized water.

2.2.2 Tween-80 solution 1%. . A 100 mL of Tween-80 solution was prepared by dissolving 1 mL of pure tween-80 in 100 mL if doubly distilled deionized water, sonicated for 15 min and diluted with deionized water when it became transparent.

2.2.3 3,4-Dihydroxybenzaldehydethiosemicarbazone (DHBTS) ($1 \times 10^{-4} \text{ mol/dm}^3$)

A 25 mL solution was prepared by dissolving 0.0528 g of recrystallised sample in dimethylformamide and it was suitably diluted to get the required concentration.

2.2.4 Mercury (II) standared solution ($4.99 \times 10^{-3} \text{ mol/dm}^3$)

A 100 mL stock solution (1 mg/mL) of divalent mercury was prepared by dissolving 135 mg of mercuric chloride (Merck, Darmstadt) in deionized water containing 1 – 2 mL of nitric acid (1 + 1). Aliquots of this solution were standardized with EDTA using XylenolOrange as an indicator. More dilute standard solution were prepared from this stock solution, as and when required. .

2.2.5 Procedure

Direct spectrophotometry

In each of set of different 10 mL volumetric flasks, 5 ml of buffer solution (pH 6.5), 1 mL of DHBTS($1 \times 10^{-3} \text{ mol/dm}^3$) and various volumes of $1 \times 10^{-6} \text{ mol/dm}^3$ mercury(II)finally added 1 mL of 1% Tween-80 and f solution were taken and made up to the mark with double distilled water. The absorbance was

measured at 375nm against the reagent blank. The calibration plot was prepared by plotting the absorbance against the amount of mercury(II).

Second order derivative spectrophotometry

For the above solutions, second order derivative spectra were recorded with a scan speed of fast (nearly 2400 nm min⁻¹); slit width of 1 nm with nine degrees freedom, in the wavelength range 390-510 nm. The derivative amplitude measured at wavelength 440 nm and plotted against amount of mercury(II) to obtain the calibration.

The calibration graph follows the straight line equation $Y = aC + b$; where C is the concentration of the solution, Y is measured absorbance or peak or valley height and a and b are constants. By substituting the corresponding experimental data substituted in the above equation, the calibration equations were calculated as $A_{375} = 0.0189C - 0.0494$ for zero order method, $A_{440} = 0.014C + 0.036$ for second order derivative method.

III. Results and Discussion

3.1 Factors affecting the absorbance

3.1.1 Absorption spectra

The absorption spectra of the mercury (II)-3,4-DHBTSC system in a micellar medium were recording using a spectrophotometer. The absorption spectra of the Hg(II)-3,4-DHBTSC is a symmetric curve with the maximum absorbance at 375nm and an average molar absorption coefficient of $3.25 \times 10^5 \text{ L mol}^{-1} \text{ cm}^{-1}$ (Fig. 1). The reagent blank exhibited negligible absorbance, despite having a wavelength in the same region. In all instances, measurements were made at 375 nm against a reagent blank.

3.1.2 Effect of surfactant: of the various surfactants [nonionic{poly oxyethylenedodecylether(Brij -35), Polyoxyethylenesorbitanmonoplamate (Tween-40), Polyoxyethylenesorbitanmonioleate(Tween-80), Triton-X-100} ; cationic[cetyltrimethylammoniumbromide(CTAB)] : and anionic { cetilpridinium chloride (CPC) , sodium dodecyl sulfate SDS } studied. Tween-80 was found to be the best surfactant for the system. In a 1%(V/L) Tween-80 medium however, the maximum absorbance was observed hence, a 1%(V/L) Tween-80 solution was used in the determination procedure.

Different volumes of 1%(V/L)Tween-80 were added to a fixed metal ion concentration, and the absorbance was measured according to the standard procedure. It was observed that at 20.05 ng/ml Hg-chelate methal.0.25-1.75 mL of 1%(V/V) Tween-80 produced a constant absorbance of the Hg-chelate(Fig. 2). A greater Excess of Tween-80 were not studied. For all subsequent measurements, 1mL 1%(V/V) Tween-80 was added.

3.1.3 Effect of Acidity

The absorbance of the complex reaches a maximum over a pH range of 6.0 to 6.8. The higher pH values may lead to hydrolysis of Hg(II) -3,4-DHBTSC system. The absorbance was at a maximum and constant when a 10 ml of solution (1 mg L^{-1} ; path length ,1) contained 1-7 mL (pH6.5) of acidic buffer at room temperature ($25 \pm 5 \text{ }^\circ\text{C}$) . Outside this range of acidity , the absorbance decreased (Fig. 3) . For all subsequent measurements 5.0 ml (pH 6.5)acidic buffer was added.

3.1.4 Effect of time

The reaction is very fast. Constant maximum absorbance was obtained just after dilution to volume, and remained strictly unaltered 24h.

3.1.5 Effect of reagent concentration:

Different molar excess of 3,4 -DHBTSC were added to a fixed metal -ion concentration, and the absorbances were measured according to the standard procedure. It was observed that at 20.05 ng/ml Hg metal(optical path length , 1 cm) reagent molar ratios 1:10 and 1:100 produced a constant absorbance of the Hg(II)-Chelate(Fig. 4) . A greater excess of the reagent was not studied. For all subsequent measurement 1mL of $1 \times 10^{-3} \text{ mol/dm}^3$ 3,4 -DHBTSC reagent was added.

3.1.6 Calibration graph (Beers Law and sensitivity)

The Calibration curve for the determination of mercury has been constructed (Fig. 5) using the optimum experimental conditions. The straight line calibration curve indicates that Beers' law is obeyed at least over a range of 5.01 to 50.14 ng of mercury per10 mL.Linear regression analysis of the calibration curve gives a correlation coefficient of 0.9998. The apparent molar absorptivity(C_{375}) calculated from the slop of regression line is $3.25 \times 10^5 \text{ L mol}^{-1} \text{ cm}^{-1}$ The Sandell's sensitivity(concentration for 0.001 absorbance unit) was found to be 0.06 ng cm^{-2} .

3.1.7 Effect of foreign ions :

The effect of various foreign ions that are generally associated with mercury(II) on the determination under optimum conditions developed was studied and the results are presented in Table 1. Cations like Ba(II), Sr(II) and U(VI) do not have any effect on the complex of Hg(II)-DHBTS, when present up to 4,700 µg, Zinc (II) and Tin (II) do not interfere in the determination of Mercury(II), even present up to 4,500µg. Pb(II) and Mn(II) can be tolerated up to 3,400µg. Fe(III), Cu(II) and Co(II) interference in the determination of Mercury(II), even when present in trace amounts

3.1.8 Composition of the Metal –Reagent complex

Jobs method of continuous variation (Fig .6) and the molar – ratio method were applied to ascertain the stoichiometric composition of the complex. A Hg -3,4-DHBTS(1:2) complex was indicated by both methods. The stability constants was determined by Jobs Method as 6.82×10^{11} .

3.2 Applications:

The Present method was successfully applied to the determination of mercury in sewage waste and spiked water samples.

3.2.1 Determination of mercury(II) in sewage water

The proposed method was applied for the determination of mercury(II) in sewage water collected in different parts in Visakhapatnam city. 10 g of the dried sample(sewage waste) was weighed and brought into solution by dry ash method. The results are shown in Table 2. The results showed that the concentration of mercury(II) is 6.8µg in sample-1, 9.2µg in sample-2, 6.7µg sample-3. The results are compared with atomic absorption spectroscopy(AAS) and they are found to be in good agreement.

3.2.2 Analysis of spiked water samples

The developed method was applied for the determination of Mercury(II) in spiked water samples. The recovery of mercury(II) from spiked water sample was examined using the general procedure, the results are given in Table 3. Showing that the developed method is applicable to analyse spiked water sample with the percentage recovery for mercury is 99.8 This method was applicable for the analysis of real water samples.

3.3 Second order derivative method

The second order derivative curve recorded (Fig. 7) for experimental solution showed the derivative amplitude were measured at 440 nm (peak) for different concentrations of Mercury(II) and plots were made between the amount of Hg(II) and the derivative amplitude. The plots were linear and obeyed Beer's law in the range 4.06-16.04 ng/mL at 440 nm respectively.

3.4 Comparison of results

The analytical characteristics of the zero and second order derivative methods in the present investigations for Mercury(II) were compared and present in Table 4. The results in this table reveal that second order derivative method is more sensitive and selective than the zero order method.

The sensitivity of the zero method in the present investigations was compared with those of some reported methods and presented in Table 5

IV. Conclusion

This method offers several interesting features such as simplicity, rapidity, and low cost besides sensitivity. The number of associated elements do not interfere in the determination. The selectivity of the reagent is also improved by the use of suitable masking agents to suppress the interference of metal ions like Fe(III), Cu(II), and Co(II). Hence the proposed method is recommended for the determination of Mercury(II) with DHBTS in presence of micells by spectrophotometric method, at minor and trace levels, besides its use for analysis of real samples such as industrial effluents.

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Table 1. Effect of Foreign ions in (30.08 ng/mL) amount of Mercury(II)

Ions	Tolerance limit(µg)
Ba ⁺² , Sr ⁺² , U ⁺⁶	4700
Zn ⁺² , Sn ⁺²	4500
Pb ⁺² , Mn ⁺²	3400
Fe ⁺³ , Cu ⁺² , Co ⁺²	3100
F ⁻ , Cl ⁻ , CH ₃ COO ⁻ ,	5500
I ⁻ , SO ₄ ⁻² , Br ⁻ , HCO ₃ ⁻ , SCN ⁻	5000

Table 2. Determination of Hg(II) in spiked water samples

S.No	Amount of metal ion added (µg)	Metal ion found		Recovery (%) Present method
		Present method*(µg)	AAS method	
1	30	29.6	29.7	99.0
2	60	59.5	59.8	99.3
3	90	88.7	89.8	97.2
4	120	119.4	119.7	99.3
5	150	149.3	149.8	99.0

*Average value of three determination

Table 3. Determination of mercury(II) in Sewage waste samples

S.No	Area of sewage waste	Metal ion found		Recovery (%) Present method
		Present method*(µg)	AAS method	
1	Steel plant area	6.7	7.0	99.3
2	HPCL area	9.5	9.8	99.3
3	Hindustan Zinc Ltd. Area	6.1	6.9	98.6

*Average value of three determinations,

Table 4. Comparison of results of Mercury(II)

Parameter	Zero order	Second derivative
Analytical wave length(nm)	375	440
Beer`s law range(ng/mL)	5.01-50.14	4.06-16.04
Angular coefficient(m)	0.0189	0.014
Y-ntercept(b)	0.494	0.036
Correlation coefficient(r)	0.9998	0.998
Standared deviation(s)	0.088	0.053

Table 5. Comparison with other methods

Reagent/Ref	λ_{max}	pH	Linear range	Molar absorptivity(L mol ⁻¹ cm ⁻¹)
Thiobenzoylacetone/[30]	345	4	0.6-12 μ g	1.7 x 10 ⁴
Variamine Blue/[31]	605	2.5-4	0.64-4.4 μ g	4 x 10 ⁴
Phenanthroline and eosin/[32]	550	4.5	0.2-1.2 μ g	8 x 10 ⁴
Thiacrown ether and Bromocresol Green/[33]	420	---	0.5-12 μ g	
Diphenylthiocarbazone/[34]	488	Acidic media	0.1-25 μ g	2.5 x 10 ⁴
HOTT/[35]	505	4-6	0.2-2 μ g	4 x 10 ⁴
Present work	375	6.5	5.01-50.4ng	3.25 x 10 ⁵

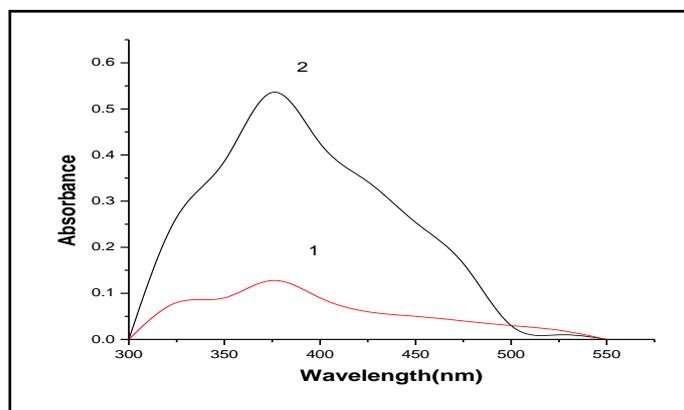


Fig. 1 1 and 2 absorption spectra of reagent blank and Hg(II)-DHBTS system ($\lambda_{max} = 375$ nm) in nonionic micellar medium

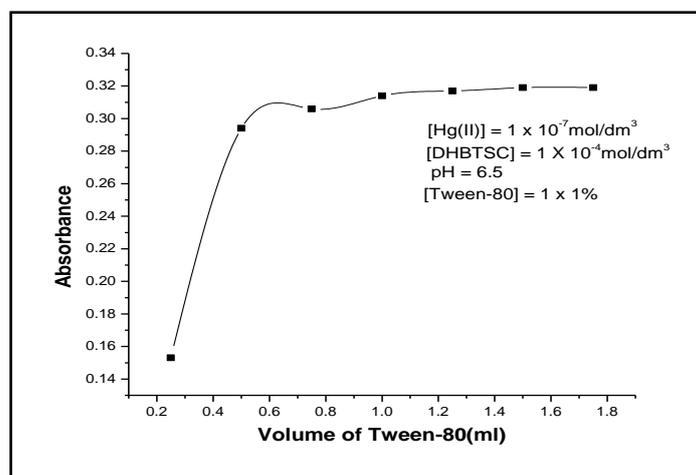


Fig. 2 Effect of surfactant on the absorbance of the Mercury(II)-DHBTS system

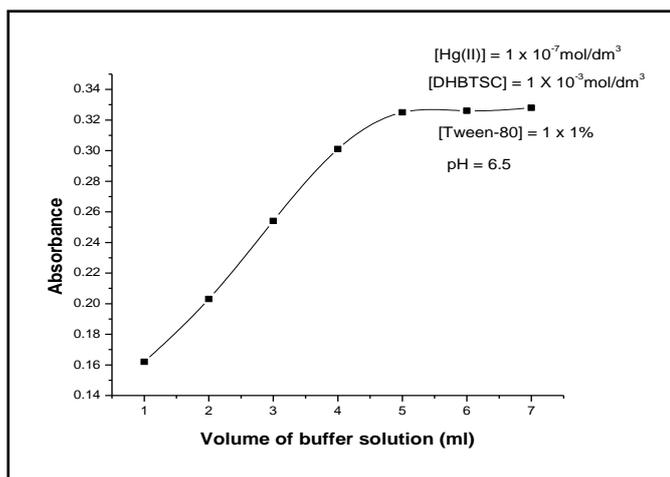


Fig. 3 Effect of Buffer solution on the absorbance of the Mercury(II)-DHBTSCT system

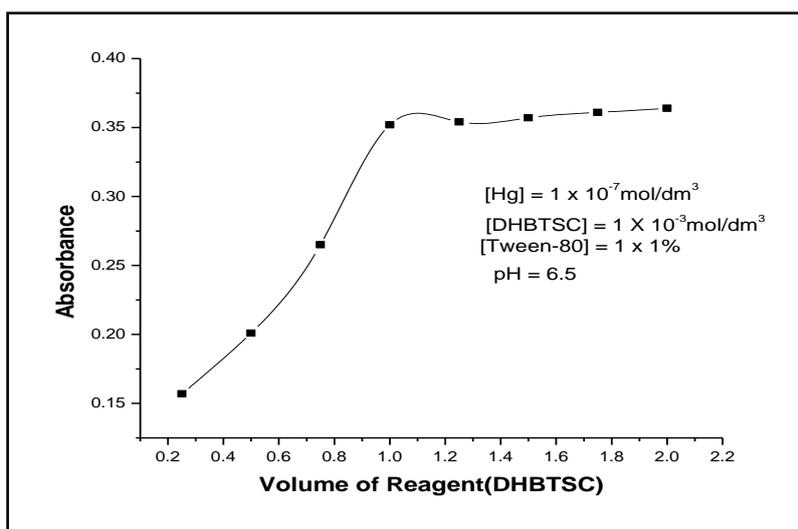


Fig. 4 Effect of reagent(DHBTSCT) on the absorbance of the Mercury(II)-DHBTSCT system.

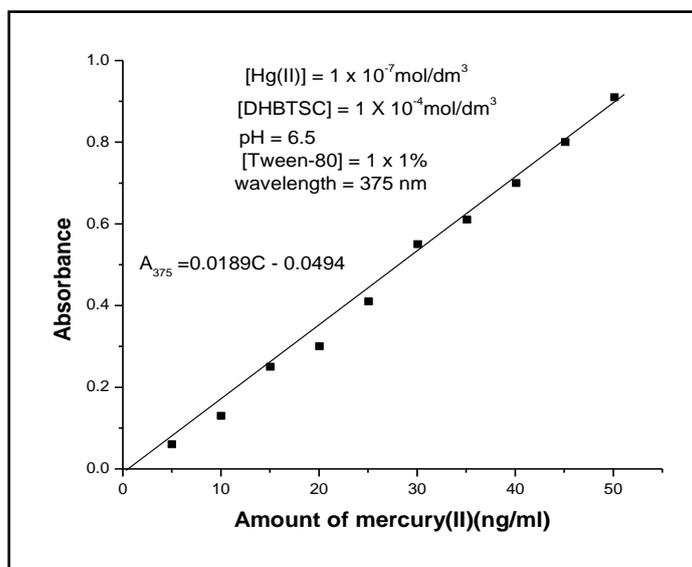


Fig. 5 Calibration graph

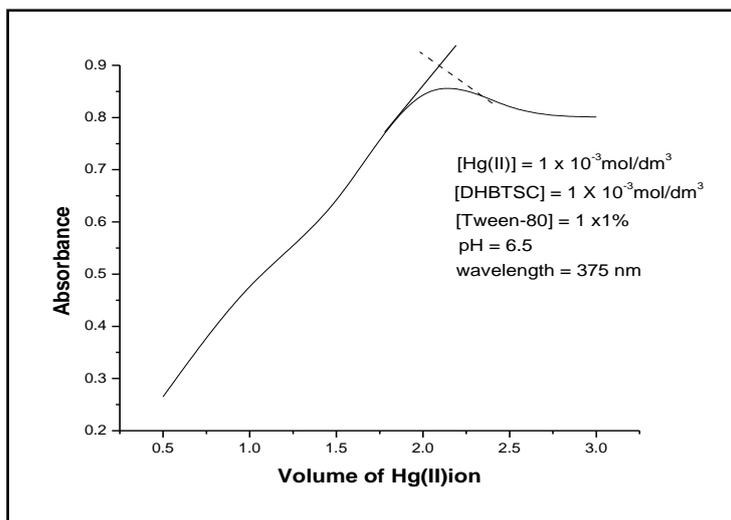


Fig. 6 Job's curve

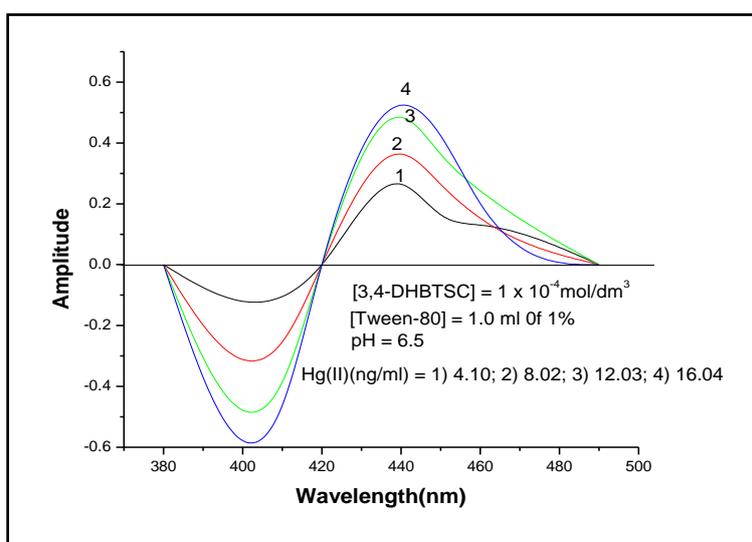


Fig. 7 Second derivative spectra of Mercury(II)-DHBTS Vs reagent blank.